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related Air Pollution in Montreal, Canada: A Case-Control Study

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Breast Cancer and Air Pollution in Montreal

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Abbreviations and Acronyms:

CI: confidence interval

NO₂: nitrogen dioxide

OR: odds ratio

PAH: polycyclic aromatic hydrocarbons

ppb: parts per billion

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Abstract:

Background. Only about 30% of cases of breast cancer can be explained by accepted risk factors. Occupational studies have shown associations between the incidence of breast cancer and exposure to contaminants found also in ambient air. We sought to determine whether the incidence of postmenopausal breast cancer is associated with exposure to urban air pollution.

Methods. We used data from a case-control study conducted in Montreal, Quebec, in 1996-1997. Cases were 383 women with incident invasive breast cancer and controls were 416 women with other incident, malignant cancers, excluding those potentially associated with selected occupational exposures. Concentrations of nitrogen dioxide (NO₂) were measured across Montreal in 2005-2006. We developed a land use regression model to predict concentrations of NO₂ across Montreal for 2006, and developed two methods to extrapolate the estimates to 1985 and 1996. We linked these estimates to addresses of residences of subjects at time of interview. We used unconditional logistic regression to adjust for accepted and suspected risk factors and occupational exposures.

Results. For each increase of 5 parts per billion of NO₂ estimated in 1996 the adjusted odds ratio was 1.31 (95% confidence interval 1.00 – 1.71). Although the size of effect varied somewhat across periods, we found an increased risk of approximately 25 percent for every increase of 5 ppb in exposure.

Conclusions. We found evidence of an association between the incidence of postmenopausal breast cancer and exposure to ambient concentrations of NO₂. Further studies are needed to confirm whether NO₂ or other components of traffic-related pollution are indeed associated with increased risks.

Introduction

Breast cancer has the highest incidence rate of all cancers in women and is the second leading cause of death from cancer in both Canada (Canadian Cancer Society 2009) and the United States (American Cancer Society 2009). Accepted risk factors for breast cancer include genetic mutations, family history of breast cancer, aspects of reproductive history, and lifestyle factors, such as alcohol consumption. Only about one-third of new cases of breast cancer are attributable to known risk factors and much of the aetiology remains unexplained (Coyle 2004). There have been consistent findings of higher rates of breast cancer in urban areas compared to rural areas, in both Canada (Bako et al. 1984) and the United States (Hall et al. 2005; Reynolds et al. 2004). Local vehicular traffic is the primary contributor to air pollution in urban areas. Vehicular emissions include gases, particles, volatile organic compounds, and polycyclic aromatic hydrocarbons (PAHs), many of which are accepted or potential carcinogens. Benzene, for example, is present in gasoline, is an accepted human carcinogen (IARC 1987) and has been shown to cause mammary carcinomas in rodents (Huff et al. 1989; Maltoni et al. 1988; Maltoni et al. 1989). Exposure to other aromatic hydrocarbons associated with gasoline, including kerosene, toluene, and xylenes have also produced increased risks of mammary cancers in female rats (Maltoni et al. 1997). Aromatic hydrocarbons are lipophilic, and may therefore reach elevated concentrations in breast tissue and promote carcinogenesis in the cells of the breast (Morris and Seifter 1992). We (Labrèche et al. 1997) postulated in the mid-1990s that exposure to organic solvents and other lipophilic toxics may cause breast cancer. Santodonato (1997) concluded that, with regard to the aetiology of human breast cancer, current scientific literature "provide persuasive

evidence for the hypothesis that certain carcinogenic PAHs produce a unique duality of pathologic effects encompassing both genotoxic and non-genotoxic components." Several studies have shown associations between the incidence of breast cancer and occupational exposure to benzene and to PAHs (Gammon et al. 2002; Petralia et al. 1999; Labrèche et al. 2010). Given that these same pollutants are present in vehicular exhaust and thus present in urban air pollution, it is plausible that traffic-related exposures may contribute to the incidence of breast cancer. A study conducted in Nassau and Suffolk Counties, New York State, (Lewis-Michl et al. 1996) suggested a possible increased risk of breast cancer among postmenopausal women living near areas characterised by relatively high traffic (adjusted odds ratio, compared to other areas, (OR) 1.29 and 95% confidence interval (CI), 0.77-2.15). In a case-control study in Erie and Niagara Counties, New York State, Bonner et al. (2005) used observations from fixed-site pollution monitors to estimate exposure to total suspended particulates and found that early life exposures to relatively high concentrations (i.e., > 140 μg/m³) were associated with an increased risk of developing postmenopausal breast cancer (OR 2.42, 95% CI 0.97-6.09) compared with exposure to relatively low concentrations (i.e., $< 84 \,\mu \text{g/m}^3$). In a second analysis from that study (Nie et al. 2007), estimates of residential exposures to benzo[a]pyrene, derived from a traffic emissions model, showed an OR of 2.57 (95% CI 1.16-5.69) for postmenopausal women exposed to higher concentrations at the time of first giving birth as compared to those exposed at lower concentrations. They found no evidence that exposures at other periods were associated with increased risk. Furthermore, in analyses stratified by smoking status, statistically significant associations were limited to lifetime non-smokers.

The purpose of the present paper was to determine whether the incidence of postmenopausal breast cancer in Montreal, Quebec, was associated with exposure to intra-urban concentrations of nitrogen dioxide (NO₂), a marker for traffic-related pollution.

Materials and Methods

Montreal is the second largest city in Canada, with the greater Montreal area having a population over 3.6 million people. Our study is restricted to the Island of Montreal and Nun's Island, which have a population of approximately 1.8 million people (Statistics Canada 2006).

Design of the Case-Control Study

We conducted a hospital-based case-control study of incident, invasive cases of postmenopausal breast cancer (Labrèche et al. 2010; Labrèche et al. 2003). The target population comprised postmenopausal women, age 50-75 years at time of diagnosis who in 1996 and 1997 were residents of the greater Montreal area. Eligible case subjects were diagnosed with primary, invasive breast cancer (International Classification of Diseases, 9th revision, code 174) that was confirmed histologically. Cases were identified from all 18 hospitals in the region that treated breast cancer, thus ensuring almost complete coverage of the target population. To minimise the potential for recall bias, control subjects had one of 32 other selected sites of incident, histologically-confirmed cancers. The controls were matched to the cases by hospital and approximately frequency-matched by age. The data were collected originally for a study examining risk of breast cancer associated with occupational exposures to selected substances, and thus selected sites of cancer (i.e., liver and intrahepatic bile duct; pancreas; lung, bronchus and trachea;

brain and central nervous system; leukemias and lymphomas) were excluded because of their possible association with occupational exposures. The controls were approximately frequency-matched to the cases by age.

One to three months after diagnosis, participants completed a structured questionnaire with content related to occupational history and other personal risk factors, including reproductive history, educational attainment, family history of breast cancer, age at menarche, smoking and alcohol consumption, body mass index, and home address (and duration of residence at that address) at time of diagnosis. Proxy respondents, who were close family members, completed a total of 75 of the questionnaires. Ethics committees at all participating hospitals and affiliated universities approved the protocol, and signed informed consent was obtained from participating subjects.

Occupational exposures were estimated using a standard methodology (Gérin et al. 1985; Siemiatycki 1990). Briefly, interviewers used a structured set of questionnaires and probed for details regarding each occupation that the subject had ever had, and a team of industrial hygienists and chemists attributed exposure to about 300 substances. For each substance, the team coded physical aspect, average duration of exposure in a working day, percent of working days exposed during the period, confidence that there was actual exposure to each agent using a 4-point ordinal scale (probably no exposure, and "low", "medium", and "high" confidence of exposure), and level of intensity. Occupational exposures to four agents that may be associated with breast cancer, from the results of other analyses of our study (Labrèche et al. 2010; Labrèche et al. 2003) were included in the present analysis: organic solvents with reactive metabolites; extremely low frequency magnetic fields; carbon monoxide; and PAHs from petroleum. Exposure indices were

computed for exposures before the age of 36 years (the period during which breast tissue may be more susceptible to exogenous insults, as female breast cells continue to develop until that age as described in detail previously (Labrèche et al. 1997)).

Neighbourhood deprivation may be a confounding factor in the association between breast cancer and air pollution because deprived populations often live in areas that are characterised by higher concentrations of air pollution (Crouse et al. 2009b; Jerrett et al. 2001). Thus, census data from 1996 were aggregated to the census tract level to describe socioeconomic characteristics of Montreal's neighbourhoods for assigning indicators of deprivation to subjects. Variables describing median household income and percentage of adults who did not complete high school were compiled for the 350 census tracts that included addresses of subjects (heretofore referred to as neighbourhood ecologic covariates).

Assessment of Exposure to Traffic-related Air Pollution

A dense sampling of ambient NO₂ was conducted in 2005 and 2006 (Crouse et al. 2009a; Crouse et al. 2009b). NO₂ is recognised as a marker of traffic-related pollution due to its co-locational association with other pollutants (Beckerman et al. 2008). The locations for the samplers were selected using a location-allocation model that placed samplers in areas likely to have high spatial variability in traffic-related pollution and high population densities (Kanaroglou et al. 2005). Samplers were deployed at 133 locations across the Island of Montreal on three occasions (spring, summer, winter) for two-week periods each time. The devices were Ogawa passive diffusion samplers that make use of triethanolamine-impregnated filters as an absorbent. Valid observations at 129 locations were obtained from all three sampling periods.

We used these observations to develop a land use regression model to predict concentrations of mean annual NO₂ for 2005-2006, at a resolution of 5 m across the Island (Crouse et al. 2009a). We modelled the natural logarithm of NO₂ on land use and traffic-related variables to generate an exposure surface. The model explained 80% of the variability in concentrations of NO₂. The residential addresses of cases and controls ascertained at the time of diagnosis (i.e., 1996) were linked to the exposure surface. Ninety-eight percent of subjects were geocoded to the x- and y-coordinates of their home address and two percent were geocoded to the centroid of the area represented by the six-character postal code, which in Montreal refers usually to a block face or to a large apartment complex.

Historical Estimates of Exposure

To account for the possibility that the spatial patterns have changed over time, we thus developed two separate but related methods (Chen et al., 2009; Chen et al., in press) to extrapolate our exposure surface back to 1996 and to 1985 (i.e., approximately the time of diagnosis and 10 years prior to this, respectively) using measurements of NO₂ from Environment Canada's National Air Pollution Surveillance network (See Supplemental Material, Figure 1).

The network in Montreal included 13 fixed-site stations that were used to measure hourly concentrations of several criteria pollutants. Our goal was to use the spatial patterns of concentrations of NO₂ collected at these stations to adjust our land use regression surface to reflect the spatial patterns of the past. Due to incomplete information, however, observations of annual mean concentrations of NO₂ were available at only nine stations for both 2006 and 1985, and at only 10 stations for both 2006 and 1996. We used inverse

distance weighting on the mean annual concentrations at these stations, in each year, to interpolate spatial surfaces. Two approaches were used:

- 1) We divided the interpolated surface describing concentrations of NO₂ in 1985 by the interpolated surface for 2006. We then multiplied our original land use regression model by this ratio to produce an extrapolated surface of estimated annual mean concentrations of NO₂ for 1985. This process was repeated with the data for 1996.
- 2) We used the predicted values from our land use regression surface in 2006 at the locations of the stations for which there were observations in 1985 (n = 10) and in 1996 (n = 12), respectively, and used these data to create new interpolated surfaces. Similar to the first method, we divided the new interpolated surface of the observed concentrations of NO_2 in 1985 by the interpolated surface that made use of predicted values (from the land use regression of 2006), which was then multiplied by the land use regression surface of 2006. Again, the process was repeated with the data for 1996.

The key differences between these methods of adjustment is that the first is based on interpolated surfaces created with *observed* measurements of NO₂ for 2006, but uses fewer data points, whereas the second is based on interpolated surfaces created with *predicted* estimates of NO₂ for 2006, but is based on a greater number of data points.

We also created two additional surfaces that described the mean estimates of NO₂ during the ten-year period between 1985 and 1996 (i.e., approximately the ten-year period prior to diagnosis). These final surfaces were created by adding each 1996 surface to the corresponding 1985 surface, and dividing the outcome by two. To assess the spatial variability between all seven exposure surfaces (i.e., land use regression for 2006, two extrapolated surfaces for 1985, two extrapolated surfaces for 1996, two surfaces for

1985-1996) 1,000 randomly generated points were sampled and Pearson correlation coefficients were estimated.

We used unconditional logistic regression to estimate odds ratios and associated 95%

Statistical Analysis

confidence intervals. In order not to lose subjects because of missing values for continuous non-occupational covariates, we used parametric smoothers (natural cubic splines) in age-adjusted logistic models to view the fitted response functions. Based on the fitted plots, we found cut-points for each covariable that defined categories such that the OR within each category was approximately constant. Subjects with missing values were assigned to a "missing information" category. Only age at diagnosis and the two neighbourhood ecologic covariates were treated as continuous variables. Fully-adjusted models included accepted and suspected risk factors for postmenopausal breast cancer (see Table 1): age at diagnosis; family history; education; ethnicity; age at bilateral oophorectomy; age at menarche; age at first full-term pregnancy; alcohol consumption and duration of hormonal replacement therapy. In addition, factors whose causal association with breast cancer is still uncertain were included as covariates: oral contraceptive use; smoking; total duration of breast-feeding; body mass index; neighbourhood ecologic covariates, and the selected occupational exposures before age 36 years. We also adjusted for the design variables proxy respondent status and for the hospital where subjects were diagnosed. Standard regression diagnostics were applied to identify possible influential subjects and to ensure that the models did not violate the assumptions of logistic regression.

We included NO₂ as a continuous, linear variable after verifying this assumption through the use of natural cubic spline functions (2-3 degrees of freedom) and visual inspection of the fitted exposure-response curves. Odds ratios are presented for each increase in exposure to NO₂ of 5 parts per billion (referred to as OR_{5ppb}) and for an increase across the inter-quartile range for each exposure period (OR_{IQR}) (the latter presented only in the Supplemental Material). The sensitivity of the results was assessed by analyses limited to those subjects for whom we had information on duration of residence and who had been residents at the same address for 10 years or more prior to diagnosis, by excluding proxy respondents, and by excluding controls with bladder cancer, as there is evidence that exposure to diesel exhaust may increase the risk of incidence of bladder cancer (Boffetta and Silverman, 2001).

Results

Exposure Surfaces

Concentrations of NO_2 decreased over time, with the highest mean value observed in 1985 (20.1 ppb) and the lowest in 2006 (11.3 ppb) (Table 2). The trend of decreasing concentrations is consistent with observations from Environment Canada's fixed-site stations (See Supplemental Material, Figure 2). The seven surfaces reflected also a narrowing of the distributions of NO_2 over time. We found positive correlations between 1,000 randomly sampled locations on the different exposure surfaces (r = 0.96 - 0.99) and between observed concentrations of annual mean concentrations of NO_2 at the fixed-site stations for 1985, 1996, and 2005 (See Supplemental Material, Table 1) suggesting that the spatial patterns of NO_2 did not vary importantly during the 20-year period between 1985 and 2006. Furthermore, the observed concentrations at the locations of the

fixed-site monitoring stations in 2006 were correlated with those in 1996 (0.89, 95% CI 0.60 - 0.97) and in 1985 (0.72, 95% CI 0.12 - 0.94).

Description of the Cases and Controls

A total of 1 631 subjects were potentially eligible for this study. Interviews were conducted among 608 cases and 667 control subjects, thus obtaining response rates of 81.1% for cases and 75.7% for controls. Of these 1 275 participants, 106 were deemed to be premenopausal and were therefore excluded; a further 79 participants had incomplete or inaccurate address information, and 291 subjects resided outside of the Island of Montreal. Therefore, these analyses included 799 subjects: 383 cases and 416 controls (Figure 1).

The most frequent sites of cancer in the 416 controls were colon (21.6%), uterus (19.0%), ovaries (9.1%), rectum (6.7%), and bladder (6.0%). We found associations for the generally accepted risk factors for postmenopausal breast cancer (i.e., family history, benign breast disease, education, age at menarche, duration of hormonal replacement therapy) (Table 2). There were essentially no differences in characteristics between the home neighbourhoods of the cases and controls (i.e., median household income and percentage of adults aged 15 years and older without high school diploma were: 34,239 CDN\$ and 34,436 CDN\$, and 32.6% and 32.8%, for cases and controls respectively).

Associations Between Postmenopausal Breast Cancer and Air Pollution

In the age-adjusted models, the OR_{5ppb} ranged from a low of 1.05 (95% CI 0.91 – 1.22) using estimates of exposure for 1985 to a high of 1.15 (95% CI 0.89 – 1.48) in 2006 (Table 3). In the fully-adjusted model using estimates of exposure from 1996, the OR_{5ppb} was 1.31 (95% CI 1.00 – 1.71). The ORs calculated per inter-quartile range were less

variable between exposure periods compared to those computed per 5 ppb, varying from 1.19 in 1985 to 1.30 in 1996 (See Supplemental Material, Table 2). The two methods of historical extrapolation produced almost identical associations with the risk of developing invasive breast cancer (Table 3; Supplemental Material, Table 2).

Sensitivity Analyses

Adjustment for individual-level educational status had the strongest effect on the size of the ORs. For example, in a reduced model using estimates of exposure for 2006, adjusted only for age and individual-level educational status, the OR_{5ppb} was 1.21 (95% CI 0.94 – 1.57). None of the other covariates included in the fully-adjusted models contributed more than 3% to the increase in the ORs.

Odds ratios were also computed for the 408 women (195 cases, 213 controls) who reported that they had been residents at the same address for at least 10 years prior to diagnosis. These ORs were slightly larger than those produced with the full study population, although, because of the smaller sample size, the confidence intervals were wider (Table 3). For example, in the fully-adjusted models, the OR_{5ppb} ranged from a low of 1.23 (95% CI 0.87 – 1.74) using estimates for 1985 to a high of 1.52 (95% CI 0.82 – 2.81) in 2006. Similar to the findings with the full dataset, the ORs calculated per interquartile range were much less variable between exposure periods, varying from 1.31 (0.82 - 2.09) to 1.35 (0.84 – 2.18) (See Supplemental Material, Table 2).

Odds ratios were computed also for subsets that excluded proxy respondents, whose interview could have elicited less accurate information (362 cases, 362 controls) and controls diagnosed with bladder cancer, a cancer linked to diesel exhaust fumes exposure

(383 cases, 392 controls). In both of these cases we found very similar associations to those found with the full study population (See Supplemental Material, Table 3).

Discussion and Conclusions

We found evidence of an association between exposure to outdoor concentrations of NO₂ and the incidence of postmenopausal breast cancer. Although the size of effect varied somewhat using estimates of exposure from different periods, we found an increased risk of approximately 25 percent for every increase of 5 ppb in exposure.

Several methodological strengths of our study suggest that the effect sizes are unbiased

Several methodological strengths of our study suggest that the effect sizes are unbiased and perhaps even conservative. Response rates among cases and controls were relatively high and there is no reason to believe that catchment areas of different hospitals would differ for cases and controls by level of exposure to traffic-related pollution, thus limiting the likelihood of selection bias. The selection of cancer patients as controls reduces the possibility of recall bias for the covariates, and concentrations of NO₂ were estimated using independent data sources. The cancer sites in the control group are not known or suspected to be associated with air pollution, with the possible exception of bladder cancer. Should, however, any of the control series cancer sites be shown subsequently to be causally associated with air pollution this would have the effect of attenuating the risks demonstrated in our study. The exclusion of bladder cancer, which has been associated with diesel exhaust fumes, did not substantially change the results. Additionally, we found that individual-level educational attainment was the only covariate to affect the estimate of risk by 5% or more.

We acknowledge that the present analysis provides only partial information on personal exposure to air pollution. Total personal exposure relates to a number of factors,

others. Two limitations of using the home address as a surrogate of exposure are related to population mobility: people do not necessarily live in the same home over the course of their lifetime nor do they spend all of their time at home. While it is true that many subjects may spend their days away from home, a study by Leech et al. (2002) found that Canadian adults spend on average ~67-68% of their time at home (indoors and outdoors combined). The finding of higher risks among subjects who lived for 10 years or more at the same address before diagnosis suggests that our risk estimates may be conservative. Lastly, given the inherent imprecision associated with geocoded addresses and other geographic data, as well as the fact that we used spatially-derived exposures as surrogates for personal exposures, the risk estimates presented here are likely subject to non-differential misclassification bias. Our results thus probably underestimate the true estimates of the relative risk of postmenopausal breast cancer associated with exposure to air pollution in this population.

Our findings should not be interpreted as meaning that NO₂ is a causal factor; it is more likely a marker of the complex mixture that is derived from combustion (Goldberg 2007). We found an association between exposures to traffic-related air pollution and the incidence of postmenopausal breast cancer in a city that by international standards is relatively unpolluted. Our findings are qualitatively similar to those reported in the two other studies that examined the hypothesis that breast cancer may be associated with exposure to air pollution (Bonner et al. 2005; Nie et al. 2007). Our results differ somewhat, however, from those studies. Among postmenopausal women, those authors found associations between early life exposures to markers of air pollution and the

incidence of breast cancer, but no associations with exposures 10 and 20 years prior to diagnosis. We could not assess associations with early life exposures, nor do we know what age periods may be critical in the induction of cancer. Our analysis of occupational exposures (Labrèche et al. 2010) suggests that exposures to some compounds before the age of 36 may be more important. If this is the case, it is possible that the risks observed here are underestimated. Studies are needed to verify whether these results represent true associations or whether they are due to chance or to undetected bias. If these associations are verified, additional studies should explore potential critical periods of exposure to air pollution in relation to the development of breast cancer.

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Table 1: Distribution of risk factors and age-adjusted odds ratios (ORs) and 95% confidence intervals (CI), postmenopausal breast cancer, Montreal, Canada, n = 799

	Case	Cases		Controls		
Variables (Reference category) ^a	n	%	n	%	OR	95% CI
Mother or sister with breast cancer (No)	210	54.8	279	67.1	1	
Yes	83	21.7	47	11.3	2.36	1.58 – 3.52
No for mother, no sisters	79	20.6	76	18.3	1.40	0.97 – 2.01
Missing information	11	2.9	14	3.4	1.13	0.50 - 2.55
Oophorectomy (Never)	276	72.1	194	46.6	1	
Only one ovary removed	27	7	30	7.2	0.61	0.35 - 1.07
Age at bilateral oophorectomy						
<44	26	6.8	34	8.2	0.55	0.32 - 0.95
45-49	19	5	15	3.6	0.89	0.44 – 1.79
50-54	16	4.2	26	6.3	0.39	0.20 - 0.76
<u>≥</u> 55	17	4.4	106	25.5	0.12	0.07 - 0.20
Missing information	2	0.5	11	2.6	0.13	0.03 - 0.59

Years of education (≤7)	107	27.9	154	37	1	
8-10	78	20.4	107	25.7	1.05	0.72 - 1.54
11-17	178	46.5	138	33.2	1.82	1.30 - 2.53
≥18	20	5.2	17	4.1	1.58	0.78 - 3.17
Ethnicity (French)	232	60.6	219	52.6	1	
English & Others	98	25.6	110	26.4	0.86	0.62 - 1.20
Jewish & Italian	53	13.8	87	20.9	0.58	0.39 - 0.86
Age at menarche (≥16)	24	6.3	39	9.4	1	
14-15	97	25.3	109	26.2	1.47	0.82 - 2.62
13	110	28.7	100	24	1.82	1.02 - 3.24
12	78	20.4	98	23.6	1.28	0.71 - 2.30
<u>≤</u> 11	74	19.3	70	16.8	1.65	0.90 - 3.02
Age at first full-term pregnancy (Never)	83	21.7	77	18.5	1	
<18	15	3.9	12	2.9	1.12	0.49 - 2.55
≥18-26	166	43.3	214	51.4	0.71	0.49 - 1.04
>26-30	59	15.4	56	13.5	1.00	0.62 - 1.62

	>30	46	12	36	8.7	1.24	0.72 - 2.12
	Pregnant but never full-term	14	3.7	21	5	0.59	0.28 - 1.24
Bı	reastfeeding, duration in weeks (0)	296	77.3	303	72.8	1	
	0-80	69	18	77	18.5	0.93	0.65 - 1.34
	>80	18	4.7	36	8.7	0.53	0.30 - 0.96
Ye	ars of oral contraception use (Never)	248	64.8	290	69.7	1	
	<1	42	11	25	6	1.78	1.04 - 3.05
	≥1	93	24.3	101	24.3	0.96	0.67 - 1.38
Но	rmone replacement therapy, months(0)	170	44.4	236	56.7	1	
	1-19	53	13.8	57	13.7	1.28	0.84 - 1.95
	20-44	29	7.6	34	8.2	1.14	0.67 – 1.96
	45-74	35	9.1	21	5	2.19	1.23 - 3.92
	75-99	16	4.2	13	3.1	1.58	0.73 - 3.39
	≥100	80	20.9	55	13.2	1.98	1.33 - 2.95
В	ody mass index (18.5-<25)	190	49.6	194	46.6	1	
	25-<30	123	32.1	133	32.0	0.97	0.71 – 1.34

>30-<35	48	12.5	51	12.3	1.00	0.64 - 1.56
≥35	15	3.9	28	6.7	0.54	0.28 - 1.06
<18.5	6	1.6	9	2.2	0.69	0.24 – 1.99
Missing information	1	0.3	1	0.2	NE	NE
Tobacco exposure (None) ^b	45	11.7	52	12.5	1	
Environmental tobacco smoke only	160	41.8	148	35.6	1.17	0.66 - 2.08
Active smoker with or without exposure						
to environmental tobacco smoke	176	46	209	50.2	1.02	0.65 – 1.61
Missing information	2	0.5	7	1.7	NE	NE
Respondent (Self)	362	94.5	362	87	1	
Proxy	21	5.5	54	13	0.41	0.24 - 0.69
Alcohol status (Never drinker) ^c	193	50.4	229	55	1	
Former drinker	50	13.1	55	13.2	1.01	0.66 – 1.57
Infrequent drinker	53	13.8	53	12.7	1.15	0.75 – 1.76
Current drinker	87	22.7	79	19	1.26	0.88 - 1.81
Benign breast disease (No)	197	51.4	333	80	1	

Yes	185	48.3	83	20	3.71	2.70 - 5.08
Missing information	1	0.3	0	0	NE	NE
Occupational exposure to solvents with						
reactive metabolites (Not exposed) ^d	338	88.3	374	89.9	1	
Non-substantial, 5 years	5	1.3	4	1	1.35	0.36 - 5.09
Substantial, 5 years	6	1.6	5	1.2	1.26	0.38 - 4.17
Exposed only at R=1	12	3.1	9	2.2	1.43	0.60 - 3.45
Other exposures	22	5.7	24	5.8	0.99	0.55 - 1.81
Occupational exposure to extremely						
low magnetic fields (Not exposed) ^d	81	21.1	102	24.5	1	
Non-substantial, 5 years	107	27.9	96	23.1	1.30	0.86 – 1.97
Substantial, 5 years	48	12.5	44	10.6	1.30	0.78 - 2.17
Exposed only at R=1	8	2.1	18	4.3	0.57	0.24 - 1.37
Other exposures	139	36.3	156	37.5	1.06	0.73 - 1.55
Occupational exposure to						
carbon monoxide (Not exposed) ^d		299	78.1	339	81.5	1

Non-substantia	l, 5 years	36	9.4	21	5.0	1.87	1.07 - 3.27	
Substantial, 5 y	/ears	0	0.0	0	0.0	NE	NE	
Exposed only a	at R=1	1	0.3	1	0.2	1.18	0.07 - 19.10	
Other exposure	es	47	12.3	55	13.2	0.94	0.62 - 1.43	
Occupational exposure to polycyclic aromatic								
hydrocarbons from	petroleum (Not exposed)	^d 354	92.4	386	92.8	1		
Non-substantia	l, 5 years	9	2.3	4	1.0	2.34	0.72 - 7.63	
Substantial, 5 y	vears	2	0.5	2	0.5	1.07	0.15 - 7.69	
Exposed only a	at R=1	2	0.5	8	1.9	0.28	0.64 - 1.22	
Other exposure	es	16	4.2	16	3.8	1.03	0.51 - 2.10	

NE = not estimated

^a Categories shown are those modelled in the analyses (see Table 3).

^b This variable was compiled with information acquired from several questions. Subjects were asked if they had smoked 100 cigarettes over the course of their lifetime, along with follow-up questions, to determine whether they were former, current, or never smokers.

Subjects were also asked whether they had ever been exposed to residential or occupational environmental tobacco smoke, along with follow-up questions concerning duration and kind of exposure.

^c The categories for this variable were determined based on information from several. Subjects were asked whether there had ever been a time in their life when they had consumed one or more drinks of beer, wine, or liquor (respectively) on a monthly basis, or on a weekly basis.

^d Substantial exposure, \geq 5 years of exposure at medium or high levels of intensity; Non-substantial exposure, <5 years of exposure at medium or high levels of intensity, but still \geq 5 years of exposure at any intensity; Exposed only at R=1, exposure only at the lowest level of confidence; Other exposures, exposures totalling less than 5 years.

Table 2: Distributions of concentrations of NO₂ (ppb) in seven different exposure surfaces, Island of Montreal, Canada

Model	Minimum	25 th	Mean	75 th	Maximum	Median
Year		percentile		percentile		
2006	4.3	9.2	11.3	12.9	37.4	10.8
1996 ^a	6.0	12.9	15.6	17.8	44.5	15.1
Mean of 1985 and 1996 ^a	6.9	14.8	17.9	20.3	55.6	17.2
1985 ^a	7.8	16.5	20.1	22.8	66.8	19.3
1996 ^b	4.9	10.3	12.7	14.7	39.5	12.2
Mean of 1985 and 1996 ^b	5.6	11.6	14.3	16.6	49.3	13.8
1985 ^b	6.2	12.7	15.8	18.3	59.2	15.1

^a Extrapolated using observed concentrations of NO₂ at each fixed-site monitoring station.

^b Extrapolated using predicted concentrations of NO₂ derived from the LUR in 2006 at each fixed-site monitoring station

 $Table \ 3: Associations \ between \ ambient \ concentrations \ of \ NO_2 \ and \ postmenopausal \ breast \ cancer, \ Island \ of \ Montreal, \ Canada$

	Full dataset.			Limited to subjects who were residents at the same						
	n = 79	= 799 (383 cases, 416 controls)			addre	address for at least 10 years prior to interview.				
						n = 408 (195 cases, 213 controls)				
Exposure	Age-a	djusted	Fully-	Fully-adjusted ^a		Age-adjusted		adjusted ^a		
surface	per 5	ppb	per 5 ppb		per 5	per 5 ppb		ppb		
	OR	95% CI	OR	95% CI	OR	95% CI	OR	95% CI		
2006	1.15	0.89-1.48	1.35	0.94-1.94	1.08	0.75-1.56	1.52	0.82-2.81		
1996 ^b	1.14	0.91-1.42	1.36	0.99-1.88	1.05	0.76-1.45	1.42	0.81-2.48		
Mean of 1996 and 1985 ^b	1.10	0.90-1.34	1.25	0.94-1.65	1.06	0.80-1.41	1.34	0.83-2.16		
1985 ^b	1.07	0.90-1.27	1.17	0.91-1.50	1.06	0.83-1.36	1.28	0.84-1.93		
1996 ^c	1.10	0.91-1.32	1.31	1.00-1.71	1.02	0.78-1.34	1.34	0.84-2.14		
Mean of 1996 and 1985 ^c	1.07	0.91-1.27	1.22	0.97-1.54	1.04	0.82-1.32	1.28	0.86-1.91		
1985°	1.05	0.91-1.22	1.16	0.94-1.42	1.04	0.85-1.29	1.23	0.87-1.74		

^a Adjusted for hospital of diagnosis, mother or sister with breast cancer, oophorectomy, years of education, ethnicity, age at menarche, age at first full-term pregnancy, breastfeeding history, oral contraceptive use, hormone replacement therapy use, body mass index, exposure to tobacco smoke, respondent/proxy status, alcohol consumption, history of benign breast disease, occupational exposures to: solvents with reactive metabolites, extremely low magnetic fields, carbon monoxide, and polycyclic aromatic hydrocarbons, and two neighbourhood ecologic covariates: median household income and percentage of adults without a high school diploma. See Table 1.

^b Extrapolated using observed concentrations of NO₂ at each fixed-site monitoring station.

^c Extrapolated using predicted concentrations of NO₂ derived from the LUR in 2006 at each fixed-site monitoring station.

Figure 1: Spatial distribution of residential addresses of breast cancer cases and control subjects, Montreal, 1996-1997

